

Atmospheric sulfuric acid leaching thermodynamics from metallurgical zinc-bearing dust sludge

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ABSTRACT

In this paper, the influence of zinc and iron reaction in sulfuric acid leaching system, the influence of the parameters on the leaching of zinc sulfate and the reaction conditions for the subsequent leachate purification were systematically studied. It is suggested that the primary zinc and iron phases in metallurgical zinc-bearing dust sludge are ZnO, ZnS, ZnFe₂O₄, Fe, Fe₂O₃ and Fe₃O₄. In Zn-Fe-H₂O system, as the concentrations of zinc and iron ions in solution increase, each component therein remains unchanged. The dominant area is subjected to change with solution pH. When the potential and pH are respectively controlled within a different scope, Zn, Fe separation proceeds at the different levels. Fe₂O₃ is difficult to be leached when sulfuric acid is used as an agent. ZnO is leached while the iron ion is inhibited by pH control. It is found by comparison of lg[C]-pH maps of Fe₂O₃ and Fe(OH)₂, Fe(OH)₃ that in the treatment of subsequent leachate, Fe²⁺ can be oxidized into Fe³⁺ by adding an proper amount of the oxidant to solution, and Fe³⁺ forms Fe(OH)₃ sediment by additive NaOH in solution. This hits the mark of iron removal. The research results have important theoretical significance for the leaching and purification process of alkali leaching-electrolysis process.

1. INTRODUCTION

China is the largest country of zinc production and consumption in the world. In 2015, China's zinc outputs reached 6.27 million tons, up 2.0% YoY, more than one-third of the world's total zinc outputs. As a powerful support for the zinc smelting industry, the zinc output has fallen well short of what's needed in domestic market since a great deal of imports are yet still required in China [1-2]. As a result, zinc-bearing dust sludge has become an important source of regenerated zinc raw materials in our country. Based on China's crude steel output of 808 million tons in 2016, that is, only last year, the output of metallurgical dust sludge in China's iron and steel industry hit upon 80 million tons or so, containing 4~16 million t zinc, 24 million tons iron. If this part of zinc resources can be recovered, this has a great significance to promote the resource conservation and sustainable development, thus improving the economic efficiency and competitiveness of enterprises [3-4].

In recent years, quite a lot of studies at home and abroad have involved the zinc-bearing dust sludge, but most of them only focused on the zinc recovery process, mainly including hydrocyclone extraction of zinc [5], zinc extraction by hydrometallurgical leaching [6-7], zinc extraction by pyrogenic process and wet-fire joint process [8-9]. Zinc leaching thermodynamics from zinc-bearing dust sludge was seldom reported. In this regard, the reduction and zinc hydrometallurgy thermodynamics have increasingly won the favor of scholars [10]. Bai Shiping [11] turned out from the analysis of blast furnace gas slime reduction thermodynamics that the reduction of blast furnace gas slime was reached in a very short time (20~40min) at a higher temperature. The reduction of each metallic oxide occurred at a relatively low

start temperature: ZnO started to reduce at a temperature of 952°C, the PbO at 281.9°C, Fe₂O₃ at 324.5°C, Fe₃O₄ at 664.2°C, and FeO at 705.5°C. Ding Zhiying et al. [12] explored the thermodynamics of zinc hydrometallurgy process using the Chemical Equilibrium Modeling Code (CEMC). The results show that the zinc-containing species are mainly ammonia and hydroxylamine complex. Zn(NH₃)₂⁴⁺ in weak alkaline solution exists in the Zn(II)-NH₃-H₂O and Zn(II)-NH₃-Cl-H₂O systems. In the Zn(II)-NH₃-Cl-H₂O system, ternary complex containing ammonia and chloride increases the solubility of zinc in the neutral solutions. There are three zinc compounds, i.e. Zn(OH)₂, Zn(OH)_{1.6}Cl_{0.4} and Zn(NH₃)₂Cl₂, according to the total ammonia, chloride and zinc concentrations. The solubility of zinc depends on three compounds. These thermodynamic diagrams illustrate the impact of ammonia, chloride and zinc concentrations on zinc solubility and provide a thermodynamic reference for zinc hydrometallurgy.

In the first part of the article, the mineral composition, chemical composition and phase analysis of zinc-bearing steel dust were carried out. Then, the available data were used to calculate the standard Gibbs free energy for each leaching reaction. Under the given conditions, Booth free energy and reaction equilibrium constant K, and finally using the thermodynamic principle and existing thermodynamic data to draw the corresponding system potential-pH diagram and lg[C]-pH map, through the thermodynamic equilibrium diagram can be quickly understood as the conditions required to extract a product or purify a solution. The research results can maximize the leaching rate of zinc and provide a reliable theoretical basis for secondary resource utilization of steel zinc dust.

2. MATERIALS AND METHODS

2.1 Nature of raw materials

The Metallurgical Zinc-Bearing Dust Sludge used in the experiment is sourced from a steel mill in Hebei Province. The

results from chemical multi-element analysis are shown in Table 1. The chemical composition analysis in Table 1 was provided by Northeastern University Testing Center. The X-ray diffraction (XRD) and mineral phase analysis of zinc-bearing dust are shown in Fig. 1 and Table 2.

Table 1. Chemical composition analysis results of the zinc-bearing dust sludge /%

Element	TFe	Zn	MgO	SiO ₂	CaO	Al ₂ O ₃	Cl	K ₂ O	C	other
Content	28.45	8.71	2.91	14.32	4.56	7.23	1.87	0.45	19.84	11.65

It can be seen from Table 1 that iron, carbon and zinc are the primary valuable elements in Zinc-bearing dust sludge, followed by silicon, aluminum and calcium. The chlorine content is 1.87%, as a catalyst which causes water corrosion, and can destroy the oxide film, soluble in water, speeding up the corrosion rate of the metal surface.

As shown in Fig. 1, the composition of the zinc-bearing dust sludge raw material is complex, contains some zinc-bearing minerals such as zinc oxide (ZnO), zinc sphalerite (ZnS), franklinite (ZnFe₂O₄), aluminum zinc sulfate (ZnSO₄), some iron mines such as hematite (Fe₂O₃), magnetite (Fe₃O₄), also some carbons and a small amount of meteorine, zeolite, calcite and other silicate minerals.

From the analysis of mineral phase in Table 2, it is known that the oxidation degree of the raw material is heavy. The main zinc minerals are ZnO, followed by ZnS and ZnFe₂O₄, of which the content of ZnS accounts for 2.51%. for this kind of minerals, a selective enrichment by mineral separation gets more difficult, and generally can be achieved by pressure or oxygen pressure acid leaching; zinc sulfate accounts for 0.45%,

relatively low as it is, in the subsequent leaching process ZnO is aimed for leaching.

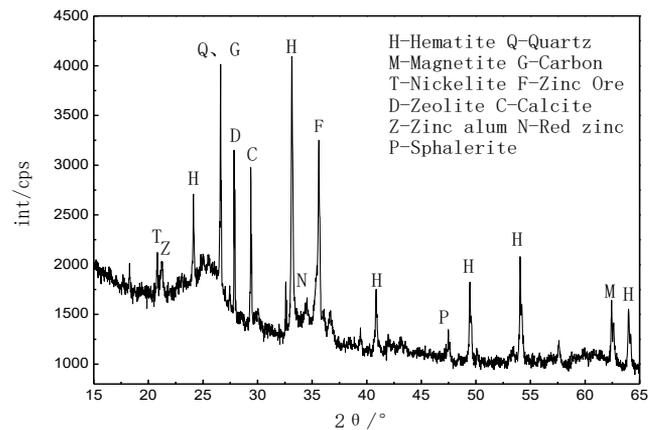
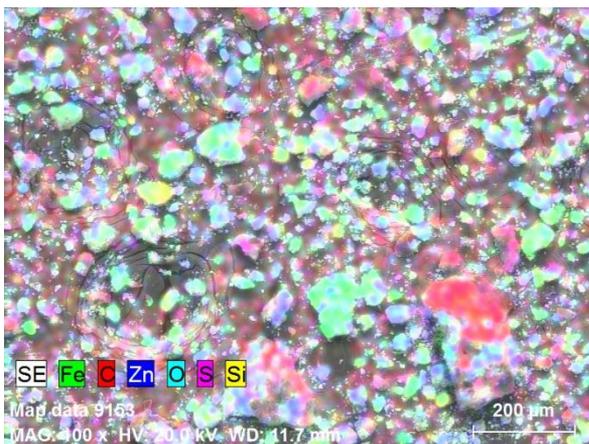


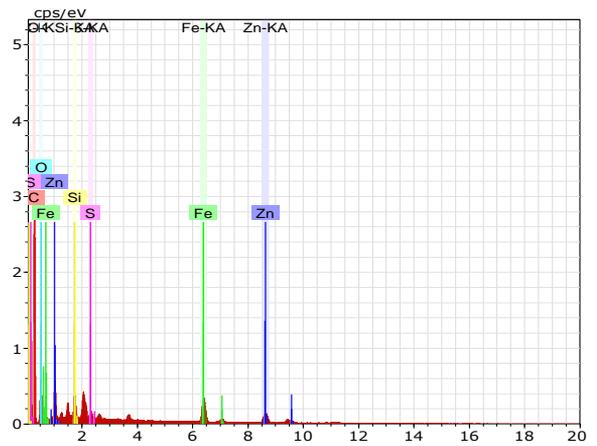
Figure 1. The XRD pattern of zinc-bearing dust sludge

Table 2. Mineral phase analysis

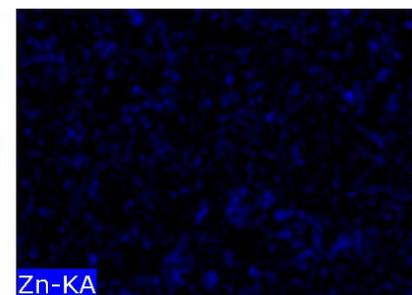
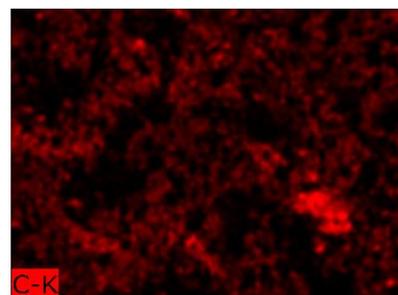
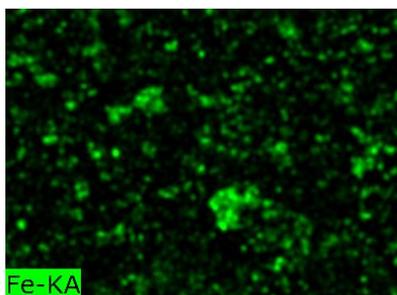
Composition	Zinc oxide	Franklinite	Zinc sphalerite	Aluminum zinc sulfate	Other zinc	TZn
Content/%	8.36	0.08	0.22	0.04	0.10	8.80
Percentage/%	95.00	0.90	2.51	0.45	1.14	100.00



(a) SEM-EDS analysis results of raw material



(b) EDS analysis results



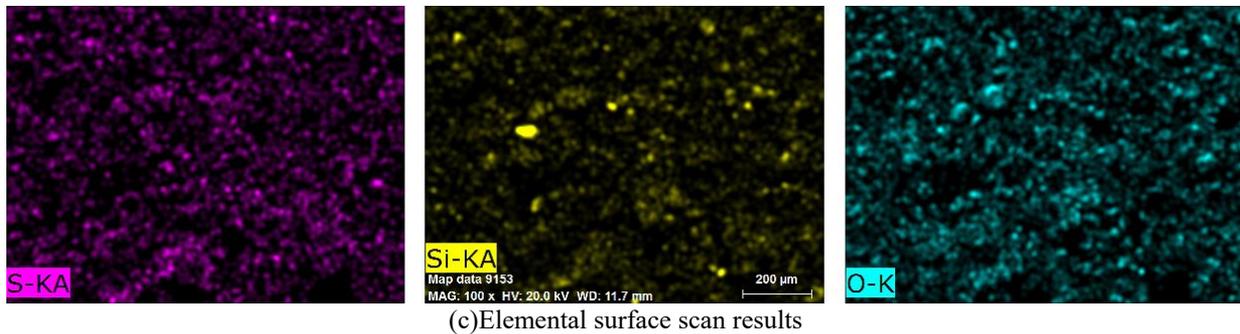


Figure 2. Area-scan of metallurgical zinc-bearing dust sludge ($\times 100$)

An area-scan was done for the distribution of various elements in metallurgical dust sludge (Fig. 2). The raw material contains Fe, C, Zn, Si, O, S and other elements, of which the contents of Fe, C, Zn elements are relatively rich. It is also known that the elements C, Si and Fe are mainly distributed in the dust sludge particles with larger diameters, whereas the elements Zn and S in the particles with smaller diameters, the elements O are distributed unevenly in the dust sludge.

2.2 Main reagents and equipment

Reagents: Sulfuric acid (AR, 98%)

Main instruments: SEM, EDS, XRD, HH-6 digital thermostat water bath, JJ-6S six-phase asynchronous electric mixer, 2XZ-1 rotary vane Type vacuum pump, blast drying oven, and pH meter.

2.3 Experimental method

The metallurgical zinc-bearing dust sludge was leached in a glassware-flask (multiple neck, 500 mL) in a water bath with temperature controller. The accuracy of leaching temperature was ± 1 °C. In each experiment, 50g metallurgical zinc-bearing

dust sludge was added to the agitated sulfuric acid solution (200 mL) of a certain concentration at a required temperature. An acid resistance mechanical agitator with stirring speed of 300 r/min was applied to agitating the reaction content. The mass ratio of liquid to solid was 4. After leaching for some time, 5 mL slurry was sampled with a pipette and immediately filtered after being diluted. The filtrate of the sample was analyzed for zinc content, and the fraction of zinc extraction was calculated for examining the leaching rate of the metallurgical zinc-bearing dust sludge.

3. LEACHING THERMODYNAMICS OF ZINC-BEARING DUST SLUDGE

3.1 Feasibility analysis of zinc sulfate thermodynamics

The thermodynamic analysis of the leaching process is made to determine the possibility and the directivity of the leaching reaction. The basic physical parameters of thermodynamics are derived from the thermodynamic data widely used in the world Library [13-15]. The reaction equation and Gibbs free energy involved in the leaching experiment are shown in Table 3.

Table 3. Reaction equation and Gibbs free energy

Number	Reaction equation	Gibbs free energy equation
(1)	$\text{ZnO(s)} + 2\text{H}^+ = \text{Zn}^{2+} + \text{H}_2\text{O(L)}$	$\Delta_r G^\theta = -91440 + 85.83\text{T(J/mol)}$
(2)	$\text{CaO(s)} + 2\text{H}^+ = \text{Ca}^{2+} + \text{H}_2\text{O(L)}$	$\Delta_r G^\theta = -193570 + 22.98\text{T(J/mol)}$
(3)	$\text{Al}_2\text{O}_3(\text{s}) + 6\text{H}^+ = 2\text{Al}^{3+} + 3\text{H}_2\text{O(L)}$	$\Delta_r G^\theta = -243800 + 484.59\text{T(J/mol)}$
(4)	$\text{MgO(s)} + 2\text{H}^+ = \text{Mg}^{2+} + \text{H}_2\text{O(L)}$	$\Delta_r G^\theta = -150980 + 96.07\text{T(J/mol)}$
(5)	$\text{CaCO}_3(\text{s}) + 2\text{H}^+ = \text{Ca}^{2+} + \text{H}_2\text{O(L)} + \text{CO}_2(\text{g})$	$\Delta_r G^\theta = -15250 - 137.53\text{T(J/mol)}$
(6)	$\text{Fe}_2\text{O}_3(\text{s}) + 6\text{H}^+ = 2\text{Fe}^{3+} + 3\text{H}_2\text{O(L)}$	$\Delta_r G^\theta = -130300 + 509.45\text{T(J/mol)}$
(7)	$\text{Fe}_3\text{O}_4(\text{s}) + 8\text{H}^+ = 2\text{Fe}^{3+} + \text{Fe}^{2+} + 3\text{H}_2\text{O(L)}$	$\Delta_r G^\theta = -211120 + 636.23\text{T(J/mol)}$
(8)	$\text{ZnFe}_2\text{O}_4(\text{s}) + 4\text{H}_2\text{SO}_4 = \text{Zn}^{2+} + \text{SO}_4^{2-} + 2\text{Fe}^{3+} + 4\text{H}_2\text{O(L)}$	$\Delta_r G^\theta = 37145730 - 393.87\text{T(J/mol)}$
(9)	$\text{ZnFe}_2\text{O}_4(\text{s}) + 4\text{H}_2\text{SO}_4 = \text{Zn}^{2+} + \text{SO}_4^{2-} + \text{Fe}_2\text{O}_3 + \text{H}_2\text{O(L)}$	$\Delta_r G^\theta = 37276030 - 115.58\text{T(J/mol)}$

The Gibbs free energy $\Delta_r G^\theta$ in the reaction is calculated based on data from thermodynamics of the existence form of the reaction, and then the $\Delta_r G^\theta - \text{T}$ of the leaching system is obtained by taking T as the abscissa and $\Delta_r G^\theta$ as the ordinate, as shown in Fig. 3.

As shown in Fig. 3, the values $\Delta_r G^\theta - \text{T}$ in reactions (1)-(5) are all less than zero at the leaching temperatures, and in reactions

(1), (2), (4), they do not change obviously with temperature build-ups; but it increases a little with temperature build-ups in reaction (3), and decreases a little in reaction (5). It is known from the calculation results that the reactants represented by reactions (1)-(5) are leached in the order of $\text{CaO} > \text{MgO} > \text{ZnO} > \text{CaCO}_3 > \text{Al}_2\text{O}_3$, while as shown in Fig. 3, the values $\Delta_r G^\theta - \text{T}$ in reactions (6)-(7) gradually increase as the temperature climbs

up, and are all greater than zero, which implies that the reactions hardly occur at room temperature, zinc can not be leached easily. It can be learned from Fig. 3 that the values $\Delta_r G_T^0$ in reactions (8)-(9) are far greater than zero, which indicates that the reactions will occur more difficultly under common leaching conditions. Consequently, the reactants

represented by reactions (6)-(9) are leached in the following order: $Fe_2O_3 > Fe_3O_4 > ZnFe_2O_4$. It further turns out that for Metallurgical Zinc-Bearing Dust Sludge, it is feasible to leach zinc at a room temperature, where a large mass of zinc can be separated with irons left in the slag.

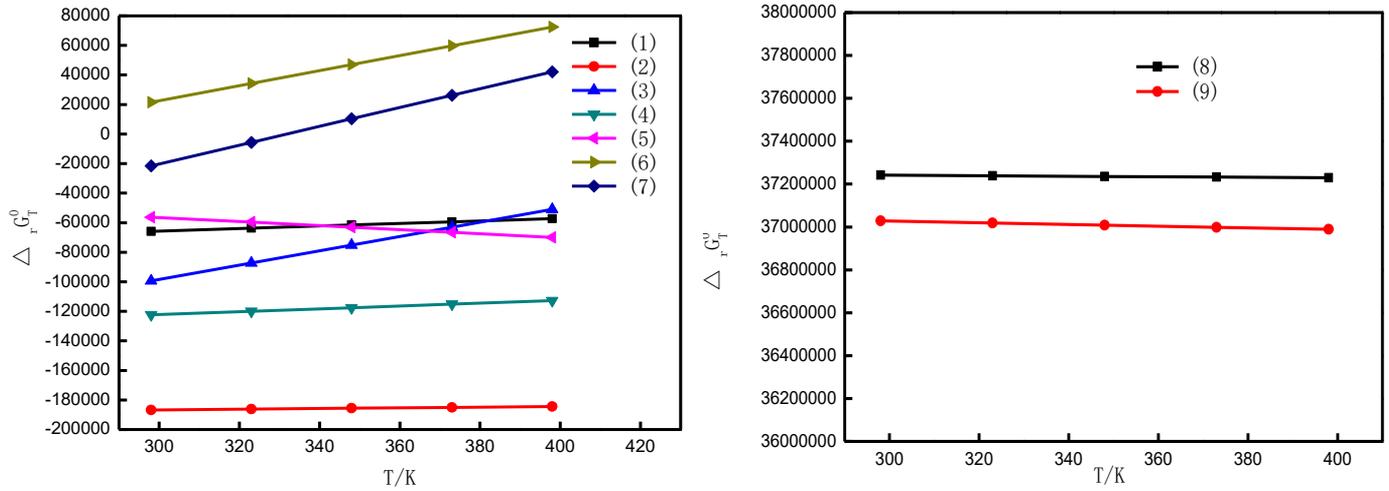


Figure 3. $\Delta_r G_T^0$ -T map for leaching system

Table 4. Zn-Fe-H₂O system potential-pH relationship

Number	Reaction equation	Potential-pH balance formula
(1)	$Fe^{3+} + e = Fe^{2+}$	$\varphi = 0.7710 + 0.0591 \lg \alpha_{Fe^{3+}} - 0.0591 \lg \alpha_{Fe^{2+}}$
(2)	$Fe_2O_3 + 6H^+ = 2Fe^{3+} + 3H_2O$	$pH = -0.2407 - 0.3333 \lg \alpha_{Fe^{3+}}$
(3)	$Fe_2O_3 + 6H^+ + 2e = 2Fe^{2+} + 3H_2O$	$\varphi = 0.7279 - 0.1773 pH - 0.0591 \lg \alpha_{Fe^{2+}}$
(4)	$ZnO \cdot Fe_2O_3 + 2H^+ = Zn^{2+} + H_2O + Fe_2O_3$	$pH = 3.3754 - 0.5000 \lg \alpha_{Zn^{2+}}$
(5)	$ZnO \cdot Fe_2O_3 + 8H^+ + 2e = Zn^{2+} + 4H_2O + 2Fe^{2+}$	$\varphi = 0.9275 - 0.2364 pH - 0.0295 \lg \alpha_{Zn^{2+}} - 0.0591 \lg \alpha_{Fe^{2+}}$
(6)	$Fe^{2+} + 2e = Fe$	$\varphi = -0.441 + 0.0295 \lg \alpha_{Fe^{2+}}$
(7)	$ZnO \cdot Fe_2O_3 + 6H^+ + 6e = ZnO + 3H_2O + 2Fe$	$\varphi = -0.0986 - 0.0591 pH$
(8)	$2H^+ + 2e = H_2$	$\varphi = -0.0591 pH$
(9)	$O_2 + 4H^+ + 4e = 2H_2O$	$\varphi = 1.2292 - 0.0591 pH$

3.2 φ -pH dominant areas of different systems

Zinc-bearing dust sludge contains some zinc and iron phases, mainly including ZnO, ZnS, ZnFe₂O₄, Fe₂O₃ and Fe₃O₄. The φ -pH dominant area maps plotted for Zn-Fe-H₂O, Zn-S-H₂O systems are computed, so do the $\lg[C]$ -pH maps for ZnO, Fe₂O₃, Fe(OH)₂, Fe(OH)₃, for the purpose of analyzing the thermodynamic behavior of sulfuric acid leaching from metallurgical zinc-bearing dust sludge.

(1) Zn-Fe-H₂O system φ -pH area maps

In Zn-Fe-H₂O system, there is a compound ZnFe₂O₄. When the temperature is 25°C, an equilibrium reaction occurs in such system on the principle of simultaneous equilibrium and electroneutrality [16-19], as shown in Fig. 4.

The φ -pH dominant area map of Zn-Fe-H₂O system is calculated by FactSage [20], as shown in Fig.4. It can be seen

from Fig. 4, as the concentrations of zinc and iron ions in the solution increase, and each component in solution does not change, the size of the dominant area is subjected to change with the solution pH. The pole potential used for the formation of target component Zn²⁺ gets lower and lower, which is more conducive to leaching. At room temperature, the control potential is greater than 0.77V, and the pH is less than -0.24, the zinc and iron will enter the solution in the forms of Zn²⁺, Fe³⁺ respectively, until the Zn²⁺, Fe³⁺ activities reach 1. When the potential and pH are controlled within II, the zinc will enter the solution selectively, while the iron remains in the slag in the form of Fe₂O₃, achieving the goal of the separation of the two while leaching. When the potential, pH are controlled within IV area, in the presence of appropriate reduction agent and pH, they can enter the solution in the forms of Zn²⁺ and Fe²⁺.

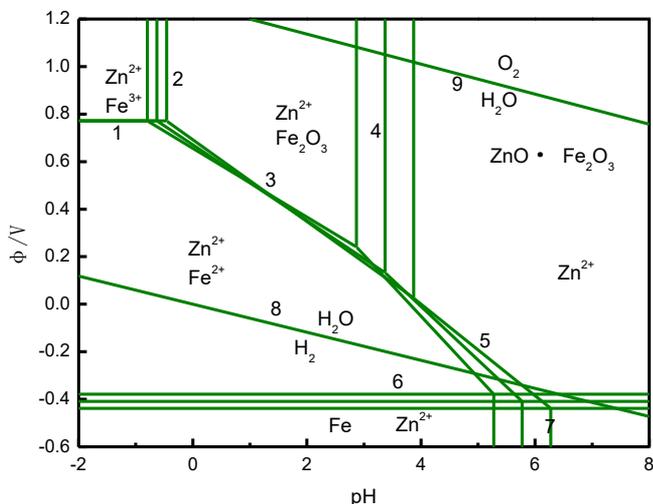


Figure 4. ϕ -pH figure of Zn-Fe-H₂O under different ion concentration of iron and zinc

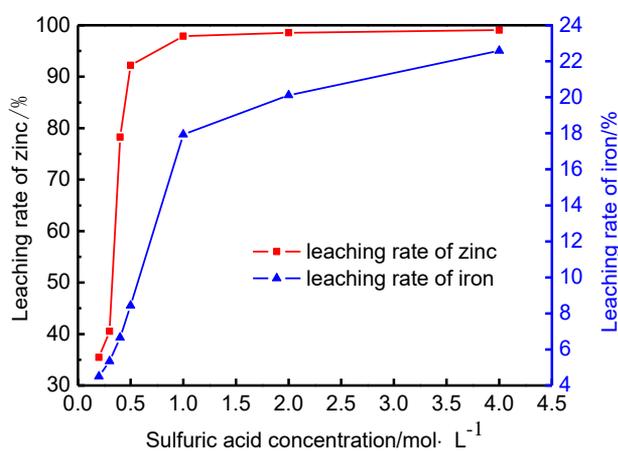


Figure 5. Leaching of the zinc-bearing dust under different sulfuric acid concentration

Figure 5 shows the effect of different sulfuric acid concentration on the leaching rate of zinc and iron in zinc-bearing dust. As can be seen from the figure, the leaching rate of zinc and iron increases with the increase of sulfuric acid

concentration. When sulfuric acid concentration is increased from 0.2mol / L increased to 0.5 mol / L, the leaching rate of zinc reached 94.00%. With the continuous increase of sulfuric acid concentration, the leaching rate of iron continued to increase, while the leaching rate of zinc increased slowly, which may be due to the formation of dissolved iron. The colloidal zinc adsorption enhanced, so that zinc failed to enter the leachate, and remain in the leaching residue, taking into account the high concentration of sulfuric acid, dissolved Fe²⁺ increase, will make subsequent leachate treatment more difficult, so you must control the sulfuric acid concentration.

(2) Potential-pH of Zn-S-H₂O system

In the Zn-S-H₂O system, there is a compound ZnS. When the temperature is 25°C, an equilibrium reaction occurs in such a system as shown in Table 5. Zn-S-H₂O system potential-pH obtained according to the reaction is shown in Fig. 6.

As learned from Fig. 6, for ZnS, the pH required for simple acid leaching is very low, i.e. -1.585. In fact, it is not feasible. For ZnS, the atmospheric oxygen-enriched leaching or the pressure acid leaching is generally adopted. When sulfuric acid is used as a leaching agent to extract zinc from zinc-bearing dust sludge, zinc sulfide gets difficult to be leached.

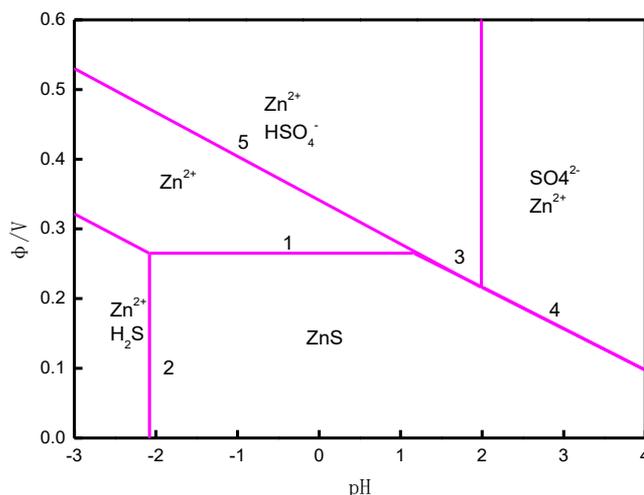


Figure 6. ϕ -pH figure of Zn-S-H₂O under different temperature(ZnS)

Table 5. Zn-S-H₂O system potential-pH relationship

Number	Reaction equation	Potential-pH balance formula
(1)	$Zn^{2+} + S + 2e = ZnS$	$\varphi = 0.265 + 0.0295 \lg \alpha_{Zn^{2+}}$
(2)	$ZnS + 2H^+ = Zn^{2+} + H_2S$	$pH = -2.08 - 0.5 \lg H_2S - 0.5 \lg \alpha_{Zn^{2+}}$
(3)	$Zn^{2+} + HSO_4^- + 7H^+ + 8e = ZnS + 4H_2O$	$\varphi = 0.320 - 0.0517 pH + 0.0074 \lg \alpha_{Zn^{2+}} + 0.0074 \lg \alpha_{HSO_4^-}$
(4)	$Zn^{2+} + SO_4^{2-} + 8H^+ + 8e = ZnS + 4H_2O$	$\varphi = 0.334 - 0.0591 pH + 0.0074 \lg \alpha_{Zn^{2+}} + 0.0074 \lg \alpha_{SO_4^{2-}}$
(5)	$Zn(OH)_2 + SO_4^{2-} + 10H^+ + 8e = ZnS + 6H_2O$	$\varphi = 0.42 - 0.074 pH + 0.0074 \lg \alpha_{SO_4^{2-}}$

3.3 lg[C]-pH map in sulfuric acid zinc-leaching system

(1) lg[C]-pH map in ZnO

Since ZnO is easily to form the hydroxyl complex ions in aqueous solution, the lg[C]-pH map of ZnO is affected to a large extent by the reactants from hydroxyl complex. There is an equilibrium as shown in Table 6, while lg[C]-pH map of

ZnO is available. C in Fig. 7 represents the total concentration of zinc ions at different pHs.

In Fig. 7, the dotted lines in purple surround a stable area for the ZnO solid phase, whose boundary lines shows the relationship of the total concentration of Zn²⁺ as a function of pHs in the system. It is known from the solubility curve of zinc oxide that the equilibrium concentration of metal ion Zn²⁺ in the solution has a tendency to first climb down and then up

with the increase of pH value since ZnO is an amphoteric oxide and easily forms hydroxyl complex ions ($Zn(OH)^+$, $Zn(OH)_3^-$, $Zn(OH)_4^{2-}$) in aqueous solution. Under neutral or weak alkaline conditions, the metals present in the form of Zn^{2+} in the solution. As pH value is on the rise, various complex ions formed with OH^- also increase. It follows that

the total concentration of metal ions in solution tends to increase when the pH value is higher. For this reason, sulfuric acid is taken as a leaching agent, the pH should be controlled within a certain range in order to facilitate the dissolution of zinc oxide, avoiding the fact that the leaching rate is affected by the generation of complex ions.

Table 6. ZnO solubility

Reaction equation	Balance equation	Equilibrium constant
$ZnO + 2H^+ = Zn^{2+} + H_2O$	$lg[Zn^{2+}] = 11.2 - 2pH$	$lgK_{S0} = 11.2$
$ZnO + H^+ = ZnOH^+$	$lg[ZnOH^+] = 2.2 - pH$	$lgK_{S1} = 2.2$
$ZnO + 2H_2O = Zn(OH)_3^- + H^+$	$lg[Zn(OH)_3^-] = -16.9 + pH$	$lgK_{S2} = -16.9$
$ZnO + 3H_2O = Zn(OH)_4^{2-} + 2H^+$	$lg[Zn(OH)_4^{2-}] = -29.7 + 2pH$	$lgK_{S3} = -29.7$

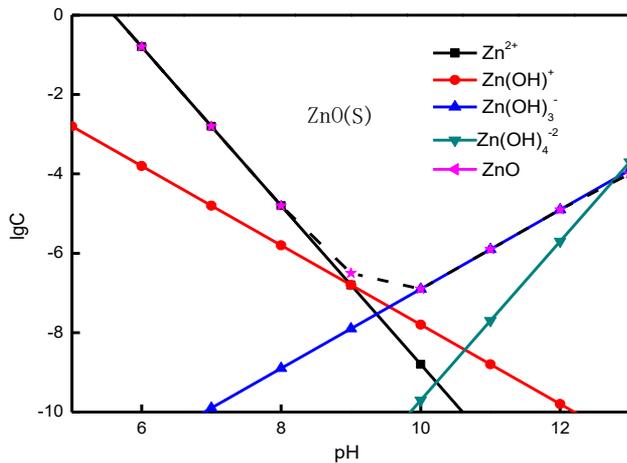


Figure 7. Solubility figure of ZnO

(2) $lg[C]$ -pH map of Fe_2O_3

Fe_2O_3 runs the following reaction in aqueous solution. From the following equilibrium, both sides take the logarithms to get the relationship between the concentration of each component and pH value, as shown in Table 7.

$lg[C]$ -pH map of Fe_2O_3 is plotted according to the equation shown in Table 7. There are several dotted yellow lines in Fig. 8 which constitute a stable area for Fe_2O_3 solid phase. This stable area has some borderlines that represent the relationship between the total iron ion concentration at different pHs in this system. It is obvious from the solubility curve of Fe_2O_3 that

the equilibrium concentration of iron ions in solution decreases first and then increases as pH value is on the rise since Fe_2O_3 is easy to form hydroxyl complex ion ($FeOH^{2+}$, $Fe(OH)_2^+$, $Fe(OH)_3$). The pH required for the dissolution of Fe_2O_3 is low even in a strong acid environment. It is therefore certain that when sulfuric acid is used as a leaching agent, it is not easy to leach Fe_2O_3 . Only when the pH is controlled in a certain scope can zinc be leached, while the leaching of iron is inhibited.

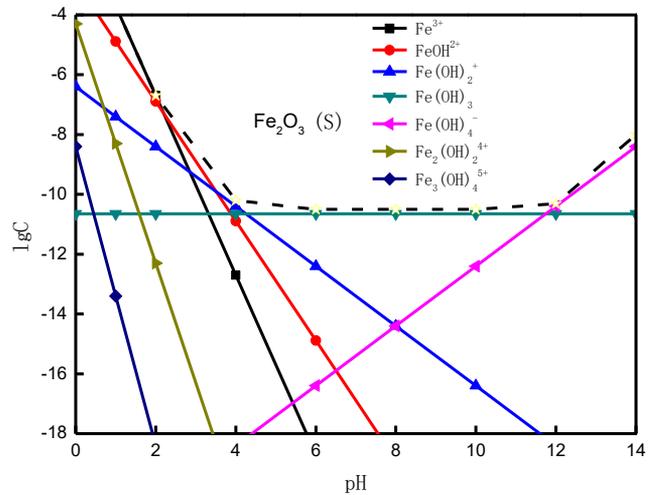


Figure 8. Solubility figure of Fe_2O_3

Table 7. Fe_2O_3 solubility

Reaction equation	Balance equation	Equilibrium constant
$\frac{1}{2}Fe_2O_3(S) + \frac{3}{2}H_2O = Fe^{3+} + 3OH^-$	$lg[Fe^{3+}] = -0.7 - 3pH$	$lg K_{sp} = -42.7$
$Fe^{3+} + OH^- = FeOH^{2+}$	$lg[FeOH^{2+}] = -2.89 - 2pH$	$lg K_1 = 11.81$
$Fe^{3+} + 2OH^- = Fe(OH)_2^+$	$lg[Fe(OH)_2^+] = -6.4 - pH$	$lg K_2 = 22.3$
$Fe^{3+} + 3OH^- = Fe(OH)_3$	$lg[Fe(OH)_3] = -10.65$	$lg K_3 = 32.05$
$Fe^{3+} + 4OH^- = Fe(OH)_4^-$	$lg[Fe(OH)_4^-] = -22.4 + pH$	$lg K_4 = 34.3$
$2Fe^{3+} + 2OH^- = Fe_2(OH)_2^{4+}$	$lg[Fe_2(OH)_2^{4+}] = -4.30 - 4pH$	$lg K_5 = 25.15$
$3Fe^{3+} + 4OH^- = Fe_3(OH)_4^{5+}$	$lg[Fe_3(OH)_4^{5+}] = -8.40 - 5pH$	$lg K_6 = 49.7$

(3) $\text{Fe}(\text{OH})_2$, $\text{Fe}(\text{OH})_3$ $\lg[C]$ -pH

When there are $\text{Fe}(\text{OH})_2$, Fe_2O_3 , $\text{Fe}(\text{OH})_3$ in solution, based on the Gibbs free energy in the standard state of each substance in the solution, an equilibrium expression between different substances is available from the reaction equation, as shown in Table 8.

The $\text{Fe}(\text{OH})_2$ $\lg[C]$ -pH is shown in Fig. 9. If $\text{pH} < 10$, Fe^{2+} appears in solution. As the pH increases, the $\text{Fe}(\text{OH})_2$ is gradually precipitated in solution. When pH value continues to increase, the precipitation of $\text{Fe}(\text{OH})_2$ is then gradually dissolved into complex ions HFeO_2^- .

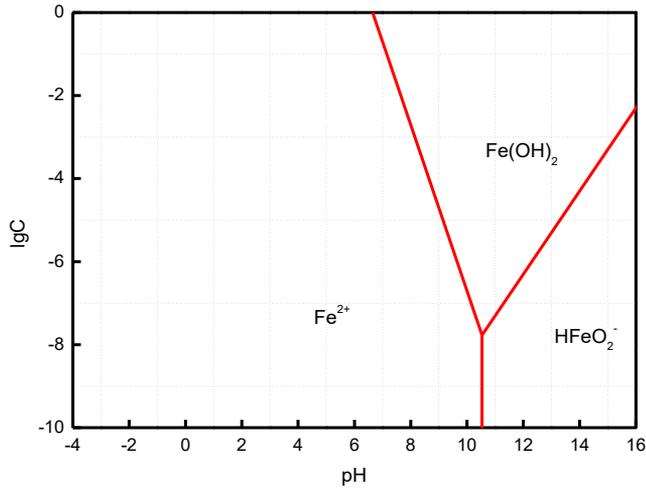


Figure 9. Solubility figure of $\text{Fe}(\text{OH})_2$ under different pH

Fe_2O_3 and $\text{Fe}(\text{OH})_3$ $\lg[C]$ -pH maps obtained according to the above equilibrium reaction are shown in Fig. 10. As can be seen, when the pH value is low, Fe^{2+} , Fe_2O_3 and complex ions FeOH^{2+} , $\text{Fe}(\text{OH})^{2+}$ may appear in solution. If $\text{pH} > 8$, stable $\text{Fe}(\text{OH})_3$ precipitation is formed in solution. In the subsequent leachate treatment, add an proper amount of oxidant such as potassium permanganate, manganese dioxide, hydrogen peroxide and the like to the solution to oxidize Fe^{2+} into Fe^{3+} , and then add NaOH to turn Fe^{3+} into $\text{Fe}(\text{OH})_3$ sediment, thus achieve the purpose of removing iron, so that the leachate has been purified, more importantly, it is more advantage for the subsequent treatment of leachate.

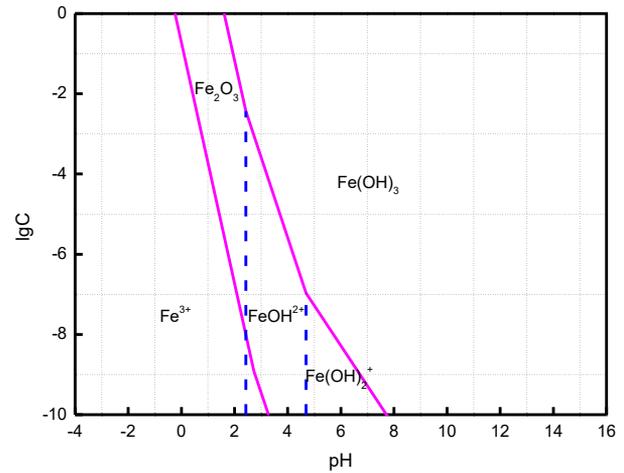


Figure 10. Solubility figure of Fe_2O_3 and $\text{Fe}(\text{OH})_3$

Table 8. $\text{Fe}(\text{OH})_2$, $\text{Fe}(\text{OH})_3$ solubility

Number	Reaction equation	Balance expression
(1)	$\text{Fe}^{2+} + \text{H}_2\text{O} = \text{FeO} + 2\text{H}^+$	$\lg[\text{Fe}^{2+}] = 13.29 - 2\text{pH}$
(2)	$\text{FeO} + \text{H}_2\text{O} = \text{HFeO}_2^- + \text{H}^+$	$\lg[\text{HFeO}_2^-] = -18.30 + \text{pH}$
(3)	$\text{Fe}^{2+} + 2\text{H}_2\text{O} = \text{HFeO}_2^- + 3\text{H}^+$	$\text{pH} = 10.53$
(4)	$2\text{Fe}^{3+} + 3\text{H}_2\text{O} = \text{Fe}_2\text{O}_3 + 6\text{H}^+$	$\lg[\text{Fe}^{3+}] = -0.72 - 3\text{pH}$
(5)	$\text{Fe}^{3+} + 3\text{H}_2\text{O} = \text{Fe}(\text{OH})_3 + 3\text{H}^+$	$\lg[\text{Fe}^{3+}] = 4.84 - 3\text{pH}$
(6)	$2\text{FeOH}^{2+} + \text{H}_2\text{O} = \text{Fe}_2\text{O}_3 + 4\text{H}^+$	$\lg[\text{FeOH}^{2+}] = -3.45 - 2\text{pH}$
(7)	$2\text{FeOH}^{2+} + \text{H}_2\text{O} = \text{Fe}(\text{OH})_3 + \text{H}^+$	$\lg[\text{FeOH}^{2+}] = 2.41 - 2\text{pH}$
(8)	$\text{Fe}^{3+} + \text{H}_2\text{O} = \text{FeOH}^{2+} + \text{H}^+$	$\text{pH} = 2.43$
(9)	$\text{FeOH}^{2+} + \text{H}_2\text{O} = \text{Fe}(\text{OH})^{2+} + \text{H}^+$	$\text{pH} = 4.69$

4. CONCLUSION

(1) In the Zn-Fe- H_2O system, the aggregation state of the dissolved components remains unchanged as the concentration of Zn^{2+} increases, but the potential of forming Zn^{2+} is gradually decreases. This is extremely conducive to the leaching reaction. In the potential-pH diagram of Fe- H_2O system, it is controlled within certain potential and pH to constitute a stable area required for leaching and leachate purification. When the potential and pH are controlled in different ranges, Zn and Fe are separated.

(2) From the ZnO, Fe_2O_3 $\lg[C]$ -pH, it can be seen that when sulfuric acid is used as the leaching agent, Fe_2O_3 is leached more difficultly. PH must be controlled within about 4 in order

to achieve an easy dissolution of zinc oxide while inhabiting Fe leaching.

(3) It has proved by comparison of $\lg[C]$ -pH maps of $\text{Fe}(\text{OH})_2$, $\text{Fe}(\text{OH})_3$ that in the subsequent leachate treatment, after adding the proper amount of oxidant to the solution to oxidize Fe^{2+} into Fe^{3+} , and then NaOH to make Fe^{3+} form $\text{Fe}(\text{OH})_3$ sediment, the iron is then removed from the solution.

(4) According to the basic principles of thermodynamics, the feasibility of leaching zinc sulfate and the stable region required for subsequent leachate purification were theoretically verified by plotting the ϕ -pH dominant region diagram and $\lg[C]$ -pH diagram of different systems. The research results have important theoretical significance for the

leaching and purification process of alkali leaching-electrolysis process.

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